

# Efficacy of Removal of CCL Viruses under Enhanced Coagulation Conditions

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The focus of coagulation as a water treatment process is shifting to accommodate recent regulatory additions that strive to balance the risks between microbial and chemical contamination of drinking water. In this work, enhanced coagulation using increased ferric chloride dose and/or pH adjustment was evaluated for removal efficacy of viruses on the United States Environmental Protection Agency (USEPA) Contaminant Candidate List (CCL), their surrogates, and dissolved organic carbon (DOC). Jar tests demonstrated that optimal DOC removal was achieved using 40 mg/L FeCl<sub>3</sub> at a pH between 5 and 6. Under these conditions, bench-scale testing resulted in a maximum removal of 2.58 log units of adenovirus type 4, 2.50 log units of feline calicivirus, 2.32 log units of MS2, 1.75 log units of PRD1, 1.52 log units of phi-X174, 2.49 log units of fr, and 56% of DOC. The trend in virus removals (MS2 and fr > PRD1 and phi-X174) was consistent between bench- and pilot-scale testing; however, pilot-plant removals exceeded bench-scale removals. Feline calicivirus was more efficiently removed than the bacteriophages, thereby suggesting potential for the bacteriophages as suitable surrogates, with MS2 and fr being more representative and PRD1 and phi-X174 (which were removed to a lesser extent) more conservative. The bacteriophages do not appear to be appropriate surrogates for adenovirus.

## Introduction

In an effort to ensure the safety of the nation's drinking water, the United States Environmental Protection Agency (USEPA) continually develops and re-evaluates regulations targeting drinking water contaminants. Individually or in tandem, recently promulgated regulations such as the Enhanced Surface Water Treatment Rule (ESWTR) and the Disinfectants and Disinfection Byproduct (D/DBP) Rule might have important implications for the water treatment industry. The ESWTR and D/DBP Rule attempt to simultaneously balance the risks posed by microbial and chemical contaminants, particularly disinfection byproducts (DBPs). The removal of DBP precursors such as dissolved organic carbon (DOC) prior to disinfection is an effective means of limiting DBP formation. Enhanced coagulation was identified by the D/DBP Rule as the best available technology to remove DOC (1).

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Traditional coagulation processes target turbidity removal, whereas enhanced coagulation targets the removal of turbidity and natural organic matter (NOM) using increased coagulant dose and/or pH adjustment (2). The Stage 1 DBP Rule requires many systems using surface water or groundwater directly under the influence of surface water to implement enhanced coagulation (3). Enhanced coagulation might become an increasingly important treatment process as water utilities adapt their systems to comply with the D/DBP Rule.

Although a great deal of attention has focused on the removal of DBP precursors from drinking water using enhanced coagulation (1, 2, 4, 5), limited information is available regarding the efficiency of enhanced coagulation for the removal of emerging pathogens (6). Conventional coagulation, flocculation, and sedimentation processes have demonstrated efficiency in the removal of a variety of microorganisms, including viruses, bacteria, and protozoa (7–9). There is a need for similar studies to determine how the optimization of enhanced coagulation for the reduction of NOM affects the removal of emerging pathogens such as those on the Contaminant Candidate List (CCL). The CCL is a list of unregulated contaminants warranting priority research because they are believed to occur in public water systems and might be associated with adverse health effects. The CCL is revised every 5 years in an effort to reflect current priorities. The second CCL, which was published in February 2005, includes 9 microorganisms and 42 chemical contaminants (10).

The overall objective of this study was to evaluate the removal of adenovirus and calicivirus (CCL viruses) and potential surrogates under enhanced coagulation conditions, as defined by DOC removal. Adenovirus type 4 and feline calicivirus, which are commonly used laboratory surrogates for enteric adenoviruses type 40 and 41 and human caliciviruses such as norovirus, were used in this study. Feline calicivirus (FCV) is not actually a CCL virus. However, it is the accepted laboratory model for studies of the removal and inactivation of human caliciviruses, for which an in vitro cell culture technique is not currently available (11, 12). Thus, FCV was used to represent the physical removal of the CCL caliciviruses throughout this study. Future advances in in vitro techniques might allow for the direct study of human calicivirus removal and inactivation, which would be of great interest.

The results of this study will aid in building a database of the efficiencies of treatments of enteric viruses and their surrogates and in evaluating CCL viruses in terms of future regulatory determination. In this study, bacteriophages representing a range of physical characteristics, as listed in Table 1, were selected as potential CCL virus surrogates. Microorganisms with removal profiles similar to those of the CCL viruses are potential surrogates for the physical removal of the CCL viruses. Surrogates would be valuable because in vitro cell culture methods, which are used in the standard assay technique for animal viruses, are labor- and time-intensive and require specialized equipment and knowledge to perform. Thus, it was important to study bacteriophages in an attempt to identify appropriate surrogate microorganisms for the CCL viruses.

## Materials and Methods

**Source Water and Coagulants.** Untreated water from the Chandler Water Treatment Plant (CWTP) in Chandler, AZ, was used as source water throughout this investigation. Influent water quality varies throughout the year based on

**TABLE 1. Characteristics of CCL Viruses and Potential Surrogate Bacteriophages**

bacteriophage	bacterial host	size (nm)	isoelectric point	genetic structure
MS2 (ATCC 15597-B1)	<i>Escherichia coli</i> (ATCC 15597)	24 <sup>13</sup> –27 <sup>14</sup>	3.5 <sup>8</sup> –3.9 <sup>13</sup>	single-stranded RNA
PRD1 (ATCC BAA-769-B1)	<i>Salmonella typhimurium</i> LT2 (ATCC 19585)	62 <sup>14</sup> –65 <sup>15</sup>	3.0 <sup>14</sup> –4.2 <sup>8</sup>	double-stranded DNA
phi-X174 (13706-B1)	<i>Escherichia coli</i> (ATCC 13706)	23 <sup>14</sup> –27 <sup>13</sup>	6.6 <sup>13</sup>	single-stranded DNA
fr (ATCC 15767-B1)	<i>Escherichia coli</i> (ATCC 19853)	23 <sup>16</sup>	8.9 <sup>8</sup> –9.0 <sup>8</sup>	single-stranded RNA

CCL virus	cell line	size (nm)	isoelectric point	genetic structure
adenovirus	PLC/PRF/5 (ATCC CRL-8024)	70 <sup>17</sup> –100 <sup>18</sup>	– <sup>a</sup>	double-stranded DNA
feline calicivirus	CRFK (ATCC CCL-94)	27 <sup>19</sup> –40 <sup>19</sup>	– <sup>a</sup>	single-stranded RNA

<sup>a</sup> Determination of the isoelectric point of pathogenic viruses is very difficult, and few have been reported. The isoelectric points of adenovirus type 4 and feline calicivirus are currently undetermined.

a fluctuating blend of surface and groundwater. Water turbidity ranged from 3 to 20 NTU, pH from 8 to 8.5, alkalinity from 125 to 200 mg/L as CaCO<sub>3</sub>, and DOC from 3 to 6 mg/L. Variations in water quality parameters have been reported to significantly impact the performance of coagulation, sedimentation, and filtration (17, 20).

Alum and ferric chloride are the most commonly used coagulants in water treatment applications (17). Ferric chloride is more efficient than alum for NOM removal (17, 21); however, there is no apparent consensus regarding coagulant selection for microbial removal. Alum was used in a related study (8), whereas ferric chloride (ferric chloride hexahydrate lumps, Sigma Chemical Co., St. Louis, MO) was used in this investigation. A cationic polymer, poly(dialyldimethylammonium chloride) (polyDADMAC; Clarifloc 350, Polydyne, Inc., Riceboro, GA), was used in conjunction with ferric chloride to improve floc settleability.

**Jar Testing.** To simultaneously optimize ferric chloride dose and pH with respect to DOC removal, a series of bench-scale jar tests were performed. All jar-test experiments were performed using a six-jar PB-700 Phipps & Bird (Richmond, VA) jar-test apparatus. In addition to DOC removal, the removals of four bacteriophages (MS2, PRD1, phi-X174, and fr), adenovirus type 4, and feline calicivirus were measured under enhanced coagulation conditions. Initial tests demonstrated that DOC concentrations were distorted by the addition of nutrient media from the stock cultures of microorganisms. Thus, separate jar tests were conducted under identical enhanced coagulation conditions for the bacteriophages, CCL viruses, and DOC. Accordingly, one jar-test experiment was conducted to develop the removal profile for the bacteriophages, which were assayed in triplicate. Two jar-test experiments were conducted using the CCL viruses because time limitations prevented duplicate assays of a single experimental sample. The two CCL tests are referred to hereafter as the “first and second CCL virus experiments”. For each of these three experiments, a separate jar-test experiment was performed to analyze DOC removal.

A 10-L sample of untreated water was collected from the CWTP for use in each jar-test experiment. The sample was transported and stored under darkened conditions at 4 °C. Jar tests were conducted within 48 h of sample collection. Prior to experimentation, the sample was allowed to return to room temperature (21 °C). The water was then seeded with a known number of microorganisms [final concentration of 1 × 10<sup>6</sup> plaque forming units (PFU) per milliliter of each

bacteriophage or 1 × 10<sup>6</sup> 50% tissue culture infective dose (TCID<sub>50</sub>) per milliliter of each virus]. The seeded sample was thoroughly mixed for 15 min before being distributed in 1.5-L portions to each of the jars. The seeded water served as the control with respect to initial virus concentrations.

Immediately following microbial seeding, ferric chloride, cationic polymer, and 1 N HCl were added to each jar. Ferric chloride doses are heavily dependent on water quality and are therefore widely variable (20). Enhanced coagulation doses range from 5 to 150 mg/L FeCl<sub>3</sub> (7), but are commonly in the range of 20–60 mg/L FeCl<sub>3</sub> (5). Coagulant doses ranging from 0 to 120 mg/L FeCl<sub>3</sub> were tested in increments of 20 mg/L in this study. Polymer was added at a constant dose of 0.4 mg/L. Variations in polymer dose have not demonstrated a significant effect on DOC removal during enhanced coagulation; therefore, polymer dose was held constant (1). The optimal pH for coagulation of NOM with ferric chloride is approximately 4–5 (2, 4); however, for practical reasons, enhanced coagulation is generally carried out in a pH range of 5–7 (1). Target pH values ranging from 5.5 to 7.5 were tested using increments of roughly 0.5 as part of this investigation.

The jar-testing protocol used throughout this study was based on the protocol described by Volk et al. (5). Immediately following chemical addition, rapid mixing was initiated at 100 rpm for 1 min. Mixing was slowed for the flocculation stage, which consisted of two 10-min mixing periods at 40 and 20 rpm. Finally, the paddles were extracted, and the contents of the jars were allowed to settle for 30 min. Samples were collected from the center of each jar, approximately 2 in. below the surface. The samples were immediately assayed for bacteriophages, CCL viruses, or DOC; pH; turbidity; and ultraviolet absorbance at a wavelength of 254 nm (UV<sub>254</sub>). UV<sub>254</sub> was used to calculate specific UV absorbance [SUVA = UV<sub>254</sub>/DOC, L (mg m)<sup>-1</sup>], which is sometimes used as a surrogate measurement of DBP precursors (14).

**Pilot-Plant Facility.** After the coagulant dose and pH had been optimized using the jar-test apparatus, pilot-plant experiments were performed at the CWTP’s pilot-plant facility using the potential surrogate bacteriophages (animal viruses were not permitted at the water treatment facility). The pilot-plant treatment train includes coagulation, flocculation, sedimentation, and filtration. A schematic is provided by Gerba et al. (8). Full-scale plant parameters such as hydraulic residence time (HRT), chemical dosing, mixing, and surface loading rates were mimicked by the small-scale pilot plant,

as shown in Table 1 in the Supporting Information. Based on full-scale facility flows, a high flow rate of 0.5 gpm (1.9 L/min) and a low flow rate of 0.25 gpm (1.0 L/min) were selected. The plant was operated for three HRTs prior to sampling such that steady-state conditions were achieved (6 and 12 h for the high and low flow rates, respectively).

The addition of bacteriophages, ferric chloride, polymer, and HCl was regulated using adjustable peristaltic pumps. Seeded water (final concentration of  $1 \times 10^6$  PFU/mL of each bacteriophage) was mixed by an inline static mixer before entering a series of three coagulation/flocculation chambers separated by diffusion walls, each equipped with a propeller-type mixer. The mixers were staged to emulate the full-scale plant tapered flocculation sequence using progressively lower mixing intensities (35 rpm in the first tank, followed by 25 and 15 rpm in the subsequent tanks). The water was then directed through the sedimentation basin before entering the dual media filters. The filters consisted of 3-in.-diameter acrylic columns containing 1.5 ft of gravel support structure, topped with 1 ft of sand, followed by 2 ft of granular activated carbon (GAC). Water samples were collected at three locations: influent, settled, and filtered. Physicochemical parameters were analyzed onsite. The samples were transported to the laboratory under darkened conditions at 4 °C and were immediately assayed for bacteriophages.

**Bacteriophage Propagation and Assays.** The bacteriophages and bacterial hosts listed in Table 1 were obtained from the American Type Culture Collection (ATCC, Rockville, MD). Bacteriophage stocks were propagated using the double agar layer (DAL) method (22). Approximately  $10^7$  PFU of bacteriophage, 1 mL of host bacteria, and 5 mL of 0.7% molten tryptic soy agar (TSA; Difco, Detroit, MI) were combined and poured onto 1.5% TSA plates. The plates were incubated at 37 °C overnight. Bacteriophage stocks were collected the following day by adding 10 mL of 1X phosphate buffer saline (PBS) to the surface of the plate and allowing it to incubate at room temperature for 1 h. The supernatant was collected and centrifuged at 4 °C at 8000g for 15 min. The pellet was discarded, and the supernatant containing the bacteriophages was stored at 4 °C. Typically, stock concentrations were on the order of  $10^{10}$  PFU/mL. New stocks were propagated once the titer dropped below  $10^8$  PFU/mL.

Immediately following jar-test and pilot-plant experiments, the water samples were serially diluted (10-fold) in 1X PBS (sufficient for up to 5-log removal) and were promptly assayed in triplicate for bacteriophages using the DAL method (22). The plates were incubated at 37 °C overnight, and plaques were counted after 12 h. Positive and negative controls were included in each set of assays for each bacteriophage.

**CCL Virus Propagation and Assays.** Adenovirus type 4 (Ad4, ATCC VR-4) was obtained from the ATCC and was cultured using primary liver cancer cells (PLC/PRF/5, ATCC CRL-8024) in 1X Eagle's minimum essential medium (MEM) containing 10% fetal bovine serum (FBS). Feline calicivirus (FCV, ATCC VR-652) was obtained from the ATCC and was cultured using Crandall feline kidney cells (CRFK, ATCC CCL-94) in 1X MEM containing 10% equine serum. The 1X MEM was supplemented with 1.5 g/L sodium bicarbonate, 15 mM Hepes, 2 mM L-glutamine, 0.1 mM non-essential amino acids, 1 mM sodium pyruvate, 100 µg/mL antimycotic, and 100 µg/mL kanamycin sulfate.

Virus stocks were propagated by inoculating the cells with a stock suspension of approximately  $1 \times 10^6$  TCID<sub>50</sub>/mL. The flasks were incubated at 37 °C until at least 90% of the cells were infected. Cells were then subjected to freeze/thaw cycles: one for Ad4 and three for FCV-infected cells (23). The supernatant was centrifuged at 4 °C at 8000g for 15 min to remove cellular debris. The virus suspension was purified and concentrated using poly(ethylene glycol) (PEG) pre-

cipitation, as described by Thurston-Enriquez et al. (23). Briefly, the suspension was augmented with 9% PEG (MW 8000) and 1 M NaCl and was allowed to stir overnight at 4 °C. It was then centrifuged at 4 °C at 8000g for 90 min. The supernatant was discarded, and the pellet was resuspended in 10% of its original volume of 1X PBS. To remove lipids and facilitate the monodispersion of viruses, a Vertrel XF (Micro Care Marketing Services, New Britain, CT) extraction was performed by centrifuging a suspension of equal parts Vertrel and virus at 4 °C at 8000g for 90 min (23). The supernatant containing the purified viruses was stored at -80 °C.

Following jar-test experiments, the water samples were serially diluted (10-fold) in 1X PBS (sufficient for up to 4.5-log removal) and were promptly assayed using conventional in vitro cell culture methodology. Cells were grown in 24-well cell culture trays. Each sample dilution was used to inoculate four wells, using 0.1 mL of sample per well. The trays were incubated in a 5% CO<sub>2</sub> incubator at 37 °C and were examined daily for any cytopathogenic effects (CPE) for up to 14 days. Typically, CPE manifested itself as cell enlargement, rounding, and the formation of grape-like clusters (24). The Karber TCID<sub>50</sub> method was used to quantify the viral concentration of each sample. The Karber TCID<sub>50</sub> equation is a statistical approach used to estimate the concentration at which 50% of the inoculated wells are positive for infection (24), as follows

$$\text{TCID}_{50} = 10^{-[\Delta - \alpha(S - 0.5)]} \quad (1)$$

where  $\Delta$  is the logarithm of the most diluted sample with 100% infectivity,  $\alpha$  is the logarithm of the dilution factor, and  $S$  is the sum of the normalized positive infections including both the last 100% and the first 0% infectious dilutions.

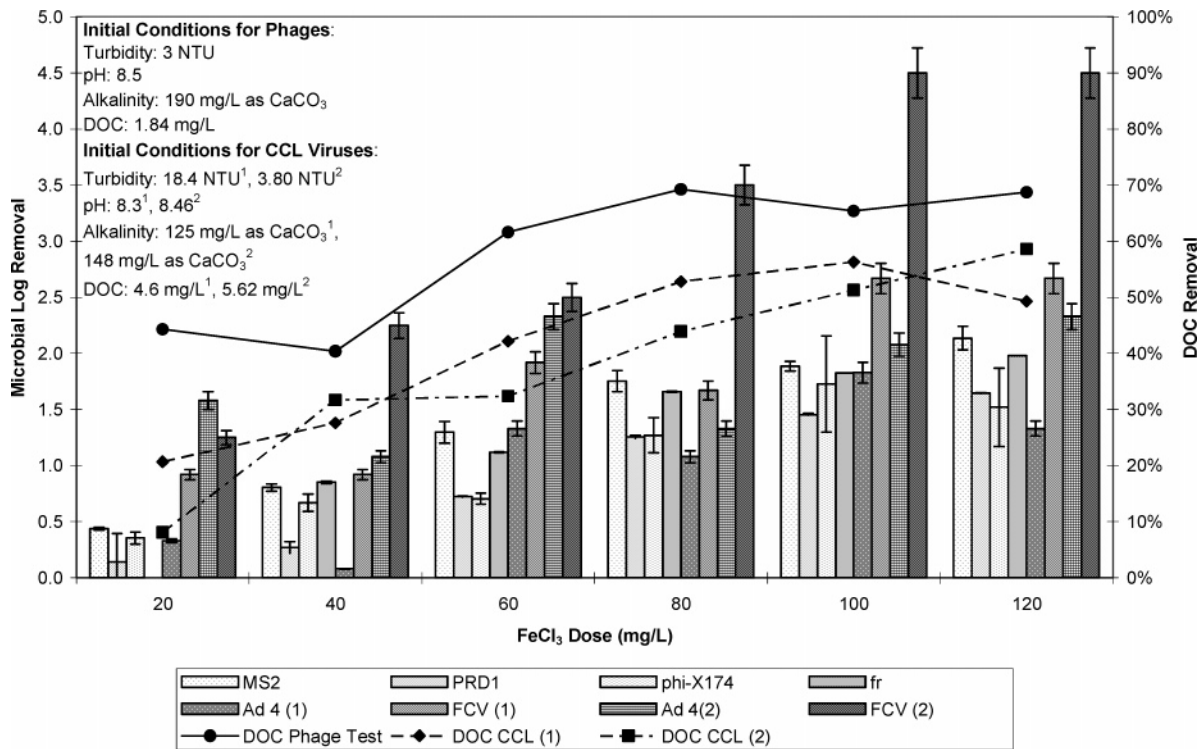
**Chemical Parameter Analysis.** The samples were analyzed for pH, turbidity, DOC, and UV<sub>254</sub>, all of which can affect the efficiency of coagulation. The pH was measured using a Mettler (Columbus, OH) pH meter. Turbidity was measured using a Hach (Loveland, CO) model 2100P turbidimeter. For DOC determination, samples were filtered using 1.2-µm Whatman (Middlesex, U.K.) GF/C glass microfiber filters, acidified using 1 N HCl, and analyzed using a Shimadzu 5050A (Kyoto, Japan) Total Organic Carbon Analyzer. A Shimadzu Multispec 1501 spectrometer was used to measure UV<sub>254</sub>.

**Statistical Analysis.** Statistical analyses were performed to determine whether the mean microbial removals resulting from different treatment conditions were statistically different at a significance level of  $\alpha = 0.05$ . SPSS (Chicago, IL) version 12.0 statistical software was used for the analyses. An analysis of variance (ANOVA) test was used to test the equality of means. The Levene statistic was computed to test the ANOVA assumption of equal variances (25). The Welch test of equality of means was used for data not satisfying the equal-variance assumption (26). Upon determination that the means were significantly different, post hoc tests were used to determine which means were different from one another. Tukey's honestly significant difference (HSD) test was used to test data with equal variances (27), and the Games-Howell test was used to analyze data that did not satisfy the equal-variances assumption (28).

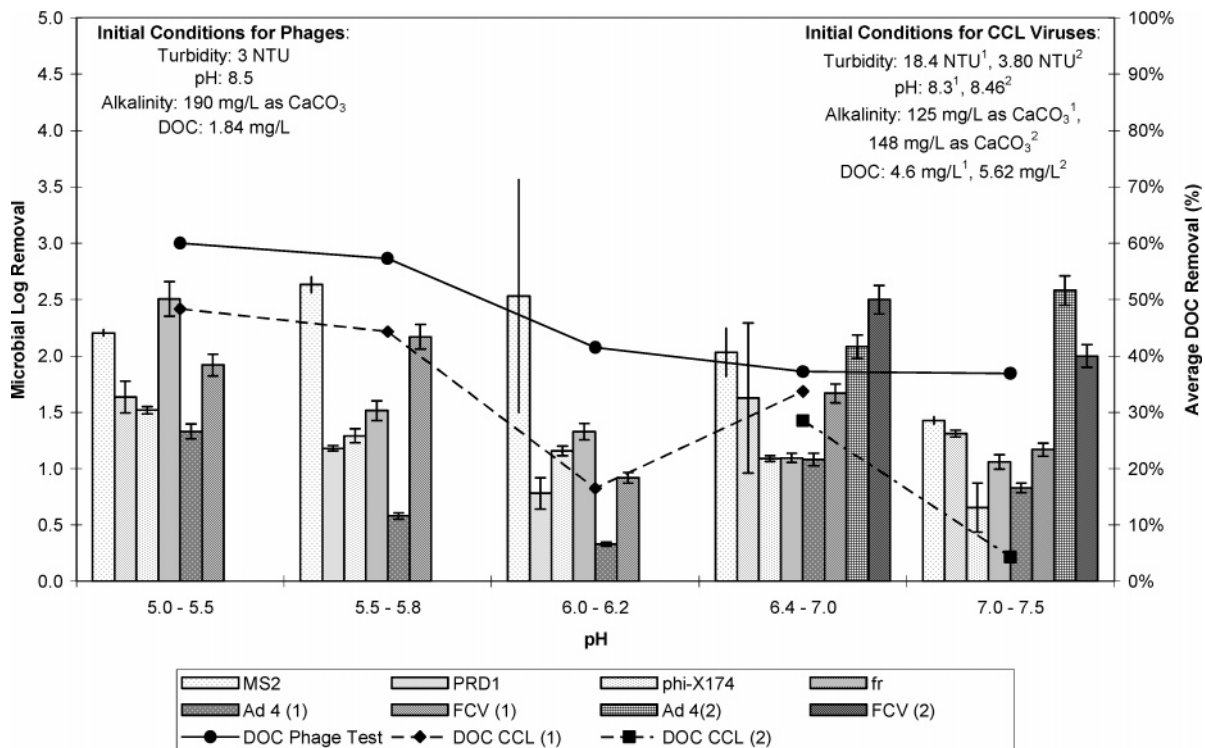
**Quality Assurance and Quality Control (QA/QC).** The standard QA/QC procedures for laboratories performing microbial and chemical analyses on water samples were followed during this research (29, 30). Additional details of the QA/QC plan are described in each subsection of the Materials and Methods.

## Results and Discussion

**Coagulant Dose-Optimization Jar Tests.** *DOC Removal.* In all three dose-optimization experiments, 25% or greater DOC



**FIGURE 1.** Removal of microbes and DOC as a function of FeCl<sub>3</sub> dose in the three jar tests: phage test, first CCL virus removal experiment (1), and second CCL virus removal experiment (2). The error bars represent ±1 standard deviation for the bacteriophages and ±5% for the CCL viruses. The DAL analysis was inconclusive for fr at a dose of 20 mg/L. Removals of FCV (1) and FCV (2) at coagulant doses of 100 and 120 mg/L FeCl<sub>3</sub> represent the maximum calculable removals based on the cell culture experiment.



**FIGURE 2.** Removal of microbes and DOC as a function of pH adjustment in the three jar tests: phage test, first CCL virus removal experiment (1), and second CCL virus removal experiment (2). The error bars represent ±1 standard deviation for the bacteriophages and ±5% for the CCL viruses. A pH range is provided because the final pH varied slightly between the jar tests. Because of difficulties adjusting the pH to target levels, no data are available for the second (2) CCL virus removal experiment for the pH range from 5 to 6.

removal was achieved using coagulant doses of 40 mg/L or greater FeCl<sub>3</sub> (Figures 1 and 2 and Tables 2–4 in the Supporting Information), which satisfies the USEPA guidelines for enhanced coagulation (3). The guidelines define

which utilities are required to practice enhanced coagulation, as dictated by initial water quality. This particular water would be exempt from the requirement based on DOC content in the source water (3), which is an extremely important

consideration for the full-scale treatment plant. However, because the objective was to evaluate virus removal resulting from enhanced coagulation, enhanced coagulation conditions were employed throughout this study, and optimization recommendations based on DOC removal were made accordingly.

The USEPA has established the optimal coagulant dose for DOC removal as the point of diminishing return (PODR) (3). The PODR is defined as the dose at which the change in DOC removal divided by the change in coagulant addition drops to (and subsequently remains below) 0.03 (3). Using the PODR criterion, the optimum ferric chloride dose for DOC removal was determined to be 40 mg/L based on the bacteriophage experiment, 80 mg/L based on the first CCL virus experiment, and 80 mg/L based on the second CCL virus experiment. The complete results of the chemical analyses for the dose-optimization experiments are provided in Tables 2–4 in the Supporting Information.

**Bacteriophage Removal.** Bacteriophage removal generally increased as coagulant dose increased (Figure 1). MS2 was generally removed to the greatest extent, followed by fr, phi-X174, and PRD1. These results are comparable to those of previous studies in that male-specific bacteriophages such as MS2 and fr were removed to a greater extent than somatic bacteriophages such as PRD1 and phi-X174 (7, 8, 31).

The quantification of virus removal during physical and chemical separation processes is well documented (32–34). It appears that the main mechanism for the physical removal of viruses is adsorption and charge neutralization followed by gravitational separation; however, the details of virus adsorption during coagulation and flocculation are not yet fully understood. Studies of transport through soil matrices have indicated that the adsorption of viruses is a complex process in which hydrophobicity, surface charge, and isoelectric point (pI) are influential factors (13, 34). It appears that no individual factor can adequately explain the mechanism of virus adsorption.

Isoelectric point has been suggested to be the dominant factor controlling virus adsorption during transport through sandy soils (13). Although the complete effects of the addition of coagulant, specifically on the surface charge of the virus, are not fully understood, the pI can help to explain adsorption during coagulation. The isoelectric points of the bacteriophages are listed in Table 1. The pI of fr is relatively high, meaning that its surface is positively charged in most naturally occurring pH ranges (6–8). This enables fr to adsorb to the surface of negatively charged particles such as  $\text{Fe}(\text{OH})_4^-$ . In the same pH range, other phages such as MS2 and PRD1 are negatively charged, thereby allowing them to adsorb to positively charged species such as  $\text{Fe}(\text{OH})_2^+$ . Theoretically, as the absolute value between pH and virus pI increases, the magnitude of the surface charge increases, thereby improving the efficacy of adsorption. This supports the observation that MS2 and fr were removed to a greater extent than phi-X174, but does not account for the relatively low removal of PRD1. Although the exact mechanisms cannot be identified as a result of this study, it is apparent that multiple factors contribute to the adsorption and resultant removal of viruses during treatment processes.

Statistical analyses were performed using the mean logarithms of the removals of the bacteriophages ( $n = 3$ ) at different coagulant doses to determine the optimal dose required to produce a significant degree of virus removal (as described in Table 6 in the Supporting Information). For each of the four bacteriophages, there was a significant difference between the log removal values at different coagulant doses ( $P \leq 0.01$ ) (Table 5 in the Supporting Information). The results indicate that 80 mg/L is the optimal dose for MS2 removal, 80 mg/L for PRD1, 40 mg/L for phi-

X174, and >120 mg/L for fr (Table 6 in the Supporting Information).

**CCL Virus Removal.** The log removal values calculated for the two CCL virus removal experiments appear to differ from one another, although this could not be verified statistically as the two experiments were not regarded as replicates because of differing initial water qualities. Overall, FCV was removed to a greater extent than Ad4 (Figure 1). There was an increase in removal of Ad4 and FCV at a dose of 60 mg/L and again at 100 mg/L for the first CCL virus experiment. For the second CCL virus experiment, there was an improvement in FCV removal at doses of 80 and 100 mg/L, whereas the results for Ad4 did not clearly identify an optimal dose. This indicates that an optimal dose for virus removal might be 60, 80, or 100 mg/L  $\text{FeCl}_3$ , which is similar to the range for the bacteriophages.

FCV appeared to be removed to a greater extent than the four bacteriophages, thereby indicating that the bacteriophages might be potential surrogates. Bacteriophages MS2 and fr would be more representative surrogates, whereas phi-X174 and PRD1 appear to be more conservative because they were removed to a lesser extent. Adenovirus was not consistently removed to a greater extent than the bacteriophages (particularly in the first CCL virus experiment); thus, it does not appear that the bacteriophages would be appropriate surrogates for Ad4.

In summary, the dose-optimization experiments suggested a wide variation in optimal coagulant doses based on different criteria (i.e., from 40 mg/L based on DOC removal to greater than 120 mg/L based on the removal of the bacteriophage fr). In light of this result, it might be beneficial to consider microbial removal in addition to DOC removal when optimizing coagulant doses. However, regulations regarding the performance of enhanced coagulation currently refer only to DOC removal. Therefore, a dose of 40 mg/L  $\text{FeCl}_3$  was identified as the optimal dose for enhanced coagulation because it was the lowest dose shown to produce a significant reduction in DOC content. Accordingly, a dose of 40 mg/L  $\text{FeCl}_3$  was used to conduct the pH-optimization jar tests and the pilot-plant study.

**pH-Optimization Jar Tests.** DOC Removal. The pH-optimization jar tests (Figure 2 and Tables 7–9 in the Supporting Information) satisfied the USEPA enhanced coagulation guidelines by achieving at least 25% DOC removal (3). The optimal pH cannot be determined using the PODR calculation, as was done to optimize dose, because the PODR is calculated using the change in DOC removal relative to coagulant dose increments. However, substantial improvements in DOC removals were observed between pH 6.01 and 5.55 and between pH 6.16 and 5.73 based on the bacteriophage and the first CCL virus removal experiments, respectively. Judging from these experiments, it appears that the optimal pH range with respect to DOC removal is less than 6.0. The complete results of the chemical analyses for the pH-optimization experiments are provided in Tables 7–9 in the Supporting Information.

**Bacteriophage Removal.** Bacteriophage MS2 was generally removed to the greatest extent, followed by fr, phi-X174, and PRD1 (Figure 2). There was a statistically significant difference between the mean log removal values at different pH values for each of the bacteriophages tested ( $P \leq 0.01$ ) (Table 10 in the Supporting Information). The post hoc tests indicate that the optimal pH is less than 6.0 for PRD1 removal and less than 5.55 for phi-X174 and fr removal, whereas an optimal pH for removal of MS2 could not be identified (Table 11 in the Supporting Information).

**CCL Virus Removal.** Differences in initial water qualities between the two tests prevented a statistical comparison of the means of the log removal values of the two CCL viruses at each pH value. However, it was apparent that FCV was

**TABLE 2. Log Removal Values of Bacteriophages at Different Stages of the Pilot Plant**

bacteriophage	low flow rate: <sup>a</sup> 0.25 gpm (5 gpm/ft <sup>2</sup> )		high flow rate: <sup>b</sup> 0.5 gpm (10 gpm/ft <sup>2</sup> )	
	settled water	filtered water	settled water	filtered water
MS2	3.94	4.60	4.28	6.87
PRD1	3.73	6.67	4.03	7.14
phi-X174	3.28	3.94	3.53	4.47
fr	3.97	4.54	4.34	7.87

<sup>a</sup> Turbidity, 13.2–17.7 NTU; pH, 8.25; alkalinity, 152 mg/L as CaCO<sub>3</sub>; DOC, 4.86 mg/L. <sup>b</sup> Turbidity, 14–19 NTU; pH, 8.2–8.25; alkalinity, 147–152 mg/L as CaCO<sub>3</sub>; DOC, 4.86 mg/L.

removed to a greater extent than Ad4. Furthermore, there was an increase in removal of viruses between pH 6.0 and 5.5 for the first CCL virus experiment (Figure 2), thereby indicating that an optimal pH for virus removal might be less than 6.0. Practical limitations effectively impart a lower pH boundary of 5 (2); thus, the recommended pH range was 5–6.

Data from the first CCL virus experiment indicate that the four bacteriophages are not appropriate CCL virus surrogates given that, in many cases, the bacteriophages were removed more efficiently. However, the results of the second CCL virus experiment show that the CCL viruses were removed to a greater extent than the bacteriophages. Thus, the bacteriophages were not eliminated as potential CCL surrogates. It would be beneficial to conduct future studies to confirm the suitability of the bacteriophages as potential CCL virus surrogates.

In summary, bench-scale testing demonstrated that the optimal conditions for enhanced coagulation were 40 mg/L FeCl<sub>3</sub> and a pH between 5 and 6. These conditions resulted in maximum removals of 2.58 log units of adenovirus type 4, 2.50 log units of feline calicivirus, 2.32 log units of MS2, 1.75 log units of PRD1, 1.52 log units of phi-X174, 2.49 log units of fr, and 56% of DOC.

**Pilot-Plant Tests.** The pilot plant was operated at a coagulant dose of 40 mg/L FeCl<sub>3</sub> and a target pH between 5 and 6 based on the optimization experiments. The pilot-scale study satisfied the USEPA's DOC removal guidelines (3) (Table 12 in the Supporting Information). The complete results of the chemical analyses for the pilot-plant experiments are provided in Table 12 in the Supporting Information.

The log removal values of the bacteriophages are listed in Table 2. Following filtration, fr was removed to the greatest extent at the high flow rate of 0.5 gpm (10 gpm/ft<sup>2</sup>), followed by PRD1, MS2, and phi-X174. At the low flow rate of 0.25 gpm (5 gpm/ft<sup>2</sup>), PRD1 was removed to the greatest extent, followed by MS2, fr, and phi-X174. The pilot-plant removals represented physical removal by coagulation, sedimentation, and filtration, whereas the jar-test data did not include filtration. Jar-test results compare to the settled water (prior to filtration) for the pilot-plant experiments. For the settled water, fr was removed to the greatest extent, followed by MS2, PRD1, and phi-X174. The trend in virus removals (MS2 and fr > phi-X174 and PRD1) was consistent between bench- and pilot-scale testing; however, pilot-plant removals exceeded bench-scale removals, which is consistent with previous reports (34).

There was a statistically significant difference between the removals of the bacteriophages at the high and low flow rates, except for the overall removal of PRD1 and the removal of fr by sedimentation ( $P \leq 0.01$ ) (Table 13 in the Supporting Information). The low flow rate should provide a better degree of removal; however, the data indicate that greater bacteriophage removals were achieved at the high flow rate. One possible explanation is the difference in initial water quality, which can affect statistical analyses; therefore, the relative efficacy of removal for the different water flow rates was not considered conclusive.

In summary, the results of the treatment optimization jar tests indicated that an optimal coagulant dose for DOC removal was 40 mg/L FeCl<sub>3</sub> with pH adjustment to between 5 and 6. Considering sedimentation only, MS2 and fr were removed to a greater extent than PRD1 and phi-X174; however, PRD1 exhibited a comparatively high degree of removal during filtration. Bacteriophages MS2, PRD1, phi-X174, and fr appear to be suitable potential surrogates for FCV, with MS2 and fr being more representative. Adenovirus type 4 seemed to be more resistant to physical and chemical treatment processes than FCV and the four bacteriophages tested, thereby indicating that the bacteriophages might not be appropriate potential surrogates for Ad4.

### Acknowledgments

This work was supported by the Office of Water, United States Environmental Protection Agency under Contract 4C-W010-NAEX. The opinions expressed in this article are those of the authors and not necessarily those of the EPA. B.K.M. is the recipient of a Department of Homeland Security Fellowship. The mention of product names does not constitute endorsement of the company.

### Supporting Information Available

Data sets for the chemical analyses and summary tables of the statistical analyses. This material is available free of charge via the Internet at <http://pubs.acs.org>.

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Received for review June 26, 2006. Revised manuscript received November 20, 2006. Accepted November 30, 2006.

ES061517Z